*Hydrological conditions predict wood density in riparian plants of south-eastern Australia*

**Abstract**

1. Wood density is a key plant functional trait which integrates the trade-offs characteristic to riparian plant ecological strategies.
2. Although dense woody tissues is costly to construct, it confers mechanical stiffness to stems, which should be useful if a plant must withstand flooding to survive. High wood density is also associated with water stress tolerance and conservative use of resources. For riparian plants, fluctuations in soil moisture driven by surface hydrology should therefore be an important driver of variation in wood density.
3. We asked the following questions in the study: (1) do riparian vegetation communities along hydrologically distinct classes of river exhibit differences in wood density? (2) is wood density related to the frequency and magnitude of flood disturbance? (3) is wood density related to predictability of water availability in the riparian zone?
4. We surveyed wood density of dominant species at fifteen riparian sites along flow-gauged rivers across south-eastern Australia. Due to the broad range of hydrological variability associated with Australian river systems, this set of sites functions as useful model for assessing the response of riparian plants to changing hydrological conditions.
5. We found wood density to vary strongly along a single axis of hydrological variability. This axis integrates flood intensity and frequency with metrics of hydrological unpredictability, and can be conceptualised as a gradient of environmental harshness.
6. *Synthesis:* Our study highlights the importance of hydrological conditions, particularly disturbance and environmental unpredictability, as determinants of ecological strategy in riparian plants. Large, rare flood events in particular appear to be driving selection for higher wood density strategies. This is likely to have significant ecological consequences for riparian plant communities in an Australian context, as well as in other regions where increasing climatic variability and frequency of extreme events are hallmarks of climate change predictions.

**Introduction**

Functional trait oriented approaches to understanding community assembly (McGill *et al.* 2006) have become increasingly common over the last decade, particularly in the plant ecological literature (Kattge *et al.* 2011). These approaches attempt to understand community assembly processes by linking morphological or physiological attributes of species to organismal success under given environmental conditions. Suites of traits can be conceptualized as axes of variation in terms of ‘ecological strategy’ and distribution of this variation across environmental gradients can provide insight into where these strategies are successful (Westoby *et al.* 2002).

Hydrology is widely considered to be the dominant abiotic force structuring riparian ecosystems. Hydrological variability in turn drives variation in moisture and substrate availability and flood disturbance, with cyclical resets to early successional conditions being characteristic of the riparian environment (Merritt *et al.* 2010). These are the conditions which are likely to dictate success of a particular plant ecological strategy in a riparian system. Several authors have recently suggested functional trait biology as a means of understanding the response of riparian plant communities to hydrological gradients (Merritt *et al.* 2010; Catford *et al.* 2012). While ecohydrological classification is becoming established as a tool to explain plant community attributes such as species richness, stand structure and composition (Poff *et al.* 2010; Arthington *et al.* 2012), functional approaches in relation to ecohydrology in riparian plant communities remain unexplored.

Woody plants determine the coarse physical structure of many riparian plant communities and are integral to the interplay of biological and physical elements that drive fluvial biogeomorphic processes (Corenblit *et al.* 2009). Consequently, an understanding of the mechanisms of riparian woody plant community assembly will provide important insights into the structure and function of fluvial landscapes. Wood density, defined as the ratio of kiln-dried mass to green volume of a wood sample (Cornelissen et al., 2003), is widely recognised as an important functional trait in plant ecology (Westoby & Wright 2006), and has been proposed as one of just several key axes of variation within which all major plant ecological strategies can be described (Westoby *et al.* 2002). Wood density is in fact an emergent property of a combination of woody tissue traits, including vessel geometry and arrangement, and the density and proportion of surrounding lignified tissue (Chave *et al.* 2009). Combined variation in these traits corresponds to the wide range of wood density strategies among woody plants.

How might different wood density strategies confer advantages to woody plant species in riparian environments? There is little direct evidence from riparian species, however general relationships between wood density and other ecological traits have been recognised in a variety of different systems and can provide some insight into the importance of variation in wood density in riparian systems. Dense wood confers mechanical stiffness (Falster 2006; Niklas & Spatz 2010), as well as resistance to pathogens (Augspurger & Kelly 1984) and herbivory (Coley 1983), but requires more investment of biomass and is therefore more costly to construct per unit of stem height. Resources not invested in structural support may be used for purposes that contribute more directly to fitness, such as production of photosynthetic tissues or propagules, or rapid growth.

According to this trade-off, it follows that several relationships between wood density and life-history strategy are apparent: wood density significantly explained variation in survival, fecundity and growth rate in a global dataset of 222 species (Adler *et al.* 2014). Studies of tropical rainforest species have shown an inverse relationship between growth rate and wood density (King *et al.* 2006; Poorter *et al.* 2008, 2010; Kraft *et al.* 2010; Wright *et al.* 2010), although no such relationship was found in a study of New Zealand tree species (Russo *et al.* 2010). Cohort survival was positively correlated with wood density in the same tropical rainforest studies (King *et al.* 2006; Poorter *et al.* 2008, 2010; Kraft *et al.* 2010; Wright *et al.* 2010). In a study of 45 rainforest species in tropical Queensland, Falster and Westoby (2005) found that wood density increased with plant height along a successional gradient. Following disturbance caused by a large cyclone in northern Queensland, Australia, wood density of rainforest trees was indicative of both damage sustained and subsequent recovery of biomass. Trees with dense wood were more likely to have experienced only minor damage. Of those trees that experienced major stem and branch damage, lower wood density trees were more likely to resprout and recover biomass faster post-disturbance (Curran *et al.* 2008). Thus it seems likely that these observed relationships in tropical rainforest communities between wood density and recovery from disturbance at the individual plant level, as well as post-disturbance succession at the community level, could also hold true for riparian systems.

While riparian plants potentially have the best access to water in a landscape, dramatic fluctuations in soil moisture are often characteristic of riparian environments. Ecological strategies for coping with intermittent water scarcity may therefore be adaptive under these conditions. The relationship between wood density and precipitation-driven patterns of soil moisture is unclear. Some studies (Weimann & Williamson, 2002; Swenson & Enquist, 2007) found little relationship between wood density and mean annual rainfall while others , found that wood density was correlated with mean annual rainfall across a transcontinental gradient (Martínez-Cabrera *et al.* 2009) and with soil moisture (Preston, Cornwell & Denoyer 2006). High wood density, along with low specific leaf area (SLA) and low maximum height, has been linked with environmental stress tolerance and conservative use of resources (Westoby 1998; Reich & Wright 2003; Swenson & Enquist 2007).Fluctuations in soil moisture driven primarily by hydrological conditions therefore should be an important driver of variation in wood density in riparian plant communities.

Given the extent to which flooding disturbance and fluctuations in water availability dominate riparian landscapes, woody tissue responses to these conditions are likely to be a primary indicator of riparian woody plant ecological strategy. Here we consider variation in wood density of dominant woody riparian plant species over a broad range of hydrological conditions, across 15 riparian sites within south-eastern Australia. As the most hydrologically variable continent on the planet (Finlayson & McMahon 1988; Peel, McMahon & Finlayson 2004) Australia provides a model system to develop insight into the possible effects of increased hydrological variability in other regions under future climates. We sought to address the following questions: (1) do riparian vegetation communities along hydrologically distinct classes of river exhibit differences in wood density? (2) is wood density related to the frequency and magnitude of flood disturbance? (3) is wood density related to predictability of water availability in the riparian zone?

**Methods**

**Study site selection**

Fifteen riparian sites were selected along gauged rivers within the South-East Coast and south-eastern Murray Darling drainage basins of Australia (Fig. 1 and Table 1). To differentiate rivers according to ecologically relevant components of hydrology, Olden and Poff (2003) described a statistical method for determining a minimally redundant set of hydrological descriptors. Kennard et al. (2010) followed this method to define a set of 120 hydrological metrics relevant to Australian rivers, which included metrics of central tendency and dispersion in all five dimensions of hydrological variation (magnitude, frequency, duration, timing, and rate of change). These metrics were then used to classify Australian river systems into twelve distinct flow regime classes, providing a foundation for analysing the properties of ecosystems across hydrological gradients. In this study, sites were drawn from rivers corresponding to ‘stable winter baseflow’, ‘unpredictable baseflow’ and ‘unpredictable intermittent’ hydrological classes, as described by Kennard et al. (2010). See Table 2 for further detail on these classes. These are the best represented hydrological classes in eastern New South Wales and Victoria, and represent clear gradients of flooding intensity and hydrological unpredictability. Five sites per hydrological class were selected based on the following criteria:

Gauged locations were selected that had >15 years of associated continuous hydrological data, and an absence of flow regulation, significant water extraction or catchment urbanisation, following Kennard et al. (2010). To minimise signals associated with human land-use and river type, the following further criteria were used to shortlist possible study sites: all were partly confined valleys with discontinuous floodplain pocket River Styles, *c.f.* Brierley & Fryirs (2005), had an intact native riparian vegetation cover (a band of native riparian vegetation extending >15 m from the bankfull channel edge), were in good geomorphic condition (lack of significant human-induced erosional or depositional landforms), minimal vegetation clearing (catchment predominantly covered by native vegetation) and occurred in a catchment smaller than 1000 km2. These criteria were assessed using a combination of visual inspection of satellite photography (Google Earth, Microsoft Bing) and information from the NSW Riparian Vegetation Extent dataset and the NSW Office of Water River Styles® geospatial dataset (Healy *et al.* 2012). To select the 15 study sites from this shortlist, accessibility by road, permission from state or private landholders, and proximity of accessible areas to continuous hydrological monitoring stations were taken into account.

**Species abundance and trait data collection**

Data collection was undertaken between December 2012 and May 2013. At each site, a 10 m by 50 m plot was marked out, with the longest edge abutting the channel edge. Criteria for selection of plot locations were: geomorphic homogeneity (the plot comprising only sloping bank where possible) and lack of anthropogenic disturbance such as built structures, roads or tracks, recent logging or clearing (in the last 20-30 years), herbicide spraying or animal grazing.

Proportional cover of woody vegetation was assessed for three strata: shrub (1-4 m), subcanopy (4-8 m) and canopy (>8 m). Species were identified using appropriate field guides, and were verified against herbarium specimens at the Macquarie University Herbarium. Some specimens were identified by staff at the Royal Botanic Gardens, Sydney.

# Wood samples were collected from dominant woody species present within the plot at >5% cover in shrub, sub canopy or canopy strata, and which had trunks robust enough to core. A 100 mm wood sample from each of two individuals per species was extracted using a 5.15 mm diameter, triple threaded increment borer (Haglöf Sweden). Samples were extracted from the base of the main trunk, 10 cm above the leaf litter level, and air-dried at 20-45 °C. On return to the laboratory, samples were rehydrated in deionised water and 10 mm sections of mature wood were cut with a razor, using visual inspection of vessel occlusion as an indicator of maturity. Sections were measured (x, y and z dimensions) with callipers (Mitzuni) to calculate wet volume, then oven-dried at 80°C for 48 hours and weighed using a microbalance (Mettler Toledo, Greifensee, Switzerland). Wood density was then calculated as the ratio of oven dry mass to wet volume (g/cm3).

**Hydrological analysis**

Daily discharge data pertaining to each field site were collated from the PINNNENA CW 10.1 database (NSW Office of Water, Department of Primary Industries) and the NSW Office of Water Continuous Water Monitoring network website (<http://realtimedata.water.nsw.gov.au/water.stm>) for NSW sites, and the Victoria State Government’s Water Measurement Information System website (<http://data.water.vic.gov.au/monitoring.htm>) for Victorian sites. Where possible 30 year time series were obtained, spanning years 1983 – 2012. Records were truncated for three sites, spanning 15, 19 and 28 years. Missing data were approximated using the Time Series Manager module in River Analysis Package (Marsh, Stewardson & Kennard 2003). Consistency of the resulting outputs were checked by visual inspection of hydrographs. For Mammy Johnson’s River, Mann River, Sportsman’s Creek and Wallagaraugh River, multiple linear regression was chosen as the most appropriate method for estimating missing data values. Linear interpolation was used for Jilliby Creek data. Daily discharge data for the remaining sites were complete.

A set of 24 hydrological metrics was pared from the full set described by Kennard et al. (2010). These metrics were chosen to be representative of variability in high flow magnitude and frequency as well as predictability and consistency of water availability in the riparian environment (see Table 3 for a description). We used the Time Series Analysis module in River Analysis Package to generate these metrics. Means and coefficients of variation were calculated for most metrics to indicate central tendency as well as spread within the data. Low and high spell metrics were thresholded at the 5th and 95th percentiles, respectively, with a flood independence criterion of 7 days between peak events. Twenty year average return interval (ARI) flood magnitude was also calculated with a flood independence value of 7 days between peaks. Colwell’s Indices were calculated using mean flow values over monthly time periods and a class distribution of 11 flow classes. Metrics of flow magnitude which had units ML / day were normalised by mean daily flow to allow for comparison between different sizes of river.

Finally, we performed permutational multivariate analysis of variance using distance matrices, using the vegan package in R (Oksanen *et al.* 2013; R Core Team 2013) across these chosen metrics to assess whether our field sites comprised three significantly different hydrological categories.

**Data analysis**

All statistical analyses were performed using the R statistical programming environment (R Core Team 2013). The R code used for these analyses can be downloaded from <https://github.com/jamesrlawson/riparian-WD/tree/master/scripts>. Statistical significance was interpreted at alpha = 0.05.

*Abundance weighted site means of wood density*

To investigate variation in wood density across hydrological gradients at the community level, abundance weighted means of wood density were generated for each site. Species relative abundance was compiled from records of % cover at the shrub (1-4 m), subcanopy (4-8 m) and canopy (>8 m) strata. Wood density values were then weighted according to species relative abundance and then summed to produce the abundance weighted site mean. This method integrates particular trait values with their real world abundance as a measure of ‘performance’, while providing a useful reduction in data dimensionality. Wood density varies only over one order of magnitude, while exhibiting relatively high intra-species plasticity. As such, abundance weighted site means work well for environmental gradient studies because the focus is maintained on the functional characteristics of the community, rather than on species *per se*.

*Comparing wood density between hydrological classes*

Raw species trait values were lumped according to the hydrological class membership, and differences between classes tested using a post-hoc Tukey’s HSD test. This test was repeated using class-lumped abundance weighted site means.

*Testing relationships between mean wood density and hydrological conditions*

Ordinary least-squares regression models were generated for selected metrics to determine relationships between hydrological gradients and abundance weighted site mean values. Wood density data was normally distributed and did not require transformation. To reduce the occurrence of Type 1 statistical error, we adjusted the resulting p values using the Benjamini - Hochberg (BH) procedure for controlling family-wise error rate (stats package, R Core Team 2013). Although ecological rationale supported inclusion of each subgroup of hydrological metrics, these metrics were highly autocorrelated. The BH procedure has been shown to control the false discovery rate for positively dependent test statistics (Benjamini and Yekutieli, 2001). We then identified ecologically relevant axes of variation in hydrological conditions by running a principal components analysis (stats package, R Core Team 2013) for the hydrological metrics that had significant relationships with site mean wood density values.

**Results**

*Confirmation of hydrological classes*

Permutational multivariate analysis of variance showed that the three hydrological classes of river were significantly different across the 24 chosen hydrological metrics. Stable winter baseflow rivers were most different from unpredictable intermittent rivers (R2 = 0.644, P= 0.011, and slightly less so from unpredictable baseflow rivers (R2 = 0.617, P = 0.005). Unpredictable intermittent and unpredictable baseflow rivers exhibited greatest similarity (R2 = 0.379, P = 0.012).

*Differences in wood density between hydrological classes*

Using abundance weighted site mean values (Fig. 2a), wood density was found to be significantly different between unpredictable baseflow rivers and stable winter baseflow rivers (P = 0.048). The difference between unpredictable intermittent rivers and stable winter baseflow rivers approached significance (P = 0.052). No significant differences in raw wood density values (Fig. 2b) were found between hydrological classes (stable winter baseflow – unpredictable baseflow P = 0.094, stable winter baseflow – unpredictable intermittent P = 0.064, unpredictable baseflow – unpredictable intermittent P = 0.994).

*How does wood density change over hydrological gradients?*

Statistics for all tests are shown in Table 4. Metrics describing the largest, most intense flood events throughout a river’s hydrological record were found to be strong positive predictors of mean wood density (Fig. 3c,d,e). Flooding frequency had no influence on wood density. Interannual variability (Fig. 3d) in flood magnitude did not show a significant relationship with wood density after Benjamini-Hochberg adjustment, although a trend was apparent. Removing the Snowy Creek site as an outlier, due to its high mean wood density (0.66 g/cm3) relative to other stable winter baseflow sites, produced a tight relationship (R2 = 75.4, P <0.001). Variability in flood rise and fall rates (Fig. 3 a,b) were also significant positive predictors of wood density, while mean flood rise and fall rates showed no significant relationship. These results indicate that large flow events not captured by the mean are driving the observed patterns of wood density.

We found denser woody tissues were increasingly favoured as baseflow index decreased (Fig. 4a). Wood density increased as patterns of average flow conditions became a) less uniformly distributed across seasons (Fig. 4c), and b) less uniformly distributed year to year (Fig. 4d). Thus mean wood density is maximised when average flow patterns are highly seasonal, but the season with which they are associated is not consistent throughout the record. A similar relationship was observed for inter-annual (Fig. 4e) but not inter-seasonal uniformity of minimum flows. In other words, it was not important how strongly minimum flows were associated with particular seasons, but whether the seasonal pattern of flows was the same across years of the record. Mean wood density also increased with increasing interannual variability in baseflow index (Fig. 4b), pointing to a strong effect from years in which baseflow deviated from the mean. Wood density also decreased with mean low spell flow (Fig. 4f), and with the mean 7 day minimum flow (after removal of Snowy Creek as an outlier) (Fig. 4g). For both metrics higher values indicate wetter minimum flow conditions. Metrics of low flow duration were not significantly associated with wood density.

*What are the principal components of variation in hydrological conditions that predict wood density?*

Hydrological metrics that significantly explained site mean wood density were highly autocorrelated in our dataset. Principal Components Analysis (PCA) identified one dominant axis within these metrics, representing 84% of variation (Fig. 5). The remaining variation was split between several minor axes, each representing less than 3% of variation.

PC1 represents a gradient of environmental harshness that integrates baseflow characteristics, seasonality and flooding intensity. Metrics which are maximised under conditions of weak seasonality, high predictability and low variability of water availability were loaded towards the positive end of the PC1 axis, while metrics that are maximised under conditions of high interannual baseflow variability and high intensity flooding were loaded towards the negative end. Stable winter baseflow rivers exhibited lower site mean wood density, and were clustered at the ‘mild’ positive end of the PC1 gradient. Unpredictable baseflow and unpredictable intermittent rivers overlapped across PC1 and were located distally towards the ‘harsh’ negative end, and were associated with higher site mean wood density. The results of the PCA analysis confirmed the pattern of differentiation in wood density between hydrological classes, and illustrated that variation in wood density is largely described by a single axis of variation from low to high variability in flow.

**Discussion**

We sought to understand how flooding disturbance and predictability of water availability might influence wood density, a key plant functional trait. To this end, we assessed the ability of hydrological classification to predict wood density of riparian plants, and then asked whether wood density could be predicted by the degree of flooding disturbance or hydrological unpredictability in the riparian zone. Our results lend credibility to broad-scale ecohydrological classification as a coarse but useful tool in understanding riparian functional ecology. Where river systems belong to different hydrological classes but are spatially or climatically close, it makes sense to dig deeper than lumped categorical comparisons. More specific insights into community assembly of riparian woody plants can be found by looking at how wood density varies along continuous hydrological gradients.

To summarise, we found evidence that mean riparian wood density is positively related to flood magnitude and extremes in flow rise and fall rates, as well as to unpredictability in flow conditions over daily, seasonal and annual timescales. Strong relationships with measures of interannual variability point to years in which the environment was extreme as powerful selectors for high wood density. Patterns of class-wise clustering were generally maintained across continua of specific hydrological gradients. Relationships were described best by quadratic or exponential models, indicating a saturation point above which variation in hydrology ceases to be associated with changes in mean wood density. We also found that for some species, intraspecific variation in wood density responded strongly to hydrology. These results were not presented here due to data limitations, but the reader is referred to the supplementary materials for further information. Removing Snowy Creek as an outlier value substantially tightened up relationships between wood density and hydrological metrics (see statistics in Table 4). This site was located within Victoria State Forestry and may have been cleared by logging within the last 20-30 years. Compared with upstream reaches within National Parks land, seral scrubs of dense stemmed *Leptospermum spp.* were considerably more abundant, which may account for this discrepancy.

Specific aspects of high flow hydrology drove variation in wood density. Wood density was not correlated with the frequency of high flow spells, which individually may not correspond to significant disturbance events, depending on the hydrological characteristics of the given river. Rather, it was the actual magnitude of flow during high flow periods that was important. The observation that variability but not average values of flood rise and fall rates was associated with wood density, indicates the influence of low frequency, intensely flashy outlier flow events not captured by the mean as important hydrological drivers. A pattern is apparent then, in which wood density in riparian communities is driven by powerful but relatively rare flow events. The abundance of high wood density strategies in these environments indicates that infrequent but high-stakes events are a greater force of selection in riparian plant communities than average conditions. This observation also suggests that selection is taking place on traits of the adult stage rather than on seedlings – indicating that persistence is more influential on population dynamics in these environments than seedling recruitment, individual growth or fecundity (Adler *et al.* 2014), or alternatively that adult and seedling traits are strongly coupled. We therefore suggest that a ‘brick house’ ecological strategy is selected for in riparian environments that experience intense flooding. This suggestion concurs with findings that trees on windy slopes tend to overcompensate for mechanical stress, with investment in defences increasing cumulatively in response to rare, extreme events (Cohen & Mangel 1999, Telewski 1995).

We can extend the observation about the influence of intense ‘pulse’ flow events on wood density: plants living in environments where flow occurs unpredictably and largely within specific events, rather than being evenly distributed throughout time, are likely to experience more intense pulses of water stress. High wood density may be symptomatic of wood anatomy strategies that allow plants to tolerate water stress (Hacke *et al.* 2001; Jacobsen & Ewers 2005; Jacobsen *et al.* 2007). Numerous studies have discussed the role of various anatomical components of woody tissue in stabilising xylem against cavitation when plants are under severe water stress, but the exact role that woody fibres play in stabilising xylem vessels appears to be inconsistent (Martínez-Cabrera *et al.* 2009). Overall, resistance against cavitation appears to result from complex interactions between wood anatomical traits (Lens *et al.* 2011; Zieminska *et al.* 2013) and/or aboveground biomass production traits, both of which are tangentially related to wood density. With the exception of ephemeral dryland rivers, riparian environments tend not to be highly water limited, so specifically constructing woody tissue to deal with constant water stress may not be advantageous. For plants that are habituated to plentiful soil moisture, however, having no backup strategy for surviving drought conditions may be risky.

A more compelling rationale for our findings is that riparian woody plants are again overcompensating for the possibility of rare, life-or-death stress events. In the absence of predictable cues about timing of watering flows, broad phenotypic plasticity in resource use traits may in fact be maladaptive in highly unpredictable environments (Valladares, Chico & Aranda 2002b; Valladares *et al.* 2002a) and conservative resource-use phenotypes such as higher wood density would be favoured (Valladares, Gianoli & Gómez 2007). Traits associated with conservative resource use and better recovery following periods of extreme stress may actually confer as much or greater fitness than traits associated with tolerating the event itself (Gutschick & BassiriRad 2003).

Conservative resource use and heavy investment in structural strength fit within the ‘resister’ category of riparian plant strategies described by Naiman & Decamps' (1997) classification of riparian plant life history strategies. ‘Invader’ strategies with which species avoid harsh hydrological conditions by achieving sexual maturity as fast as possible are also common to the riparian environment (Naiman & Decamps 1997; Woolfrey & Ladd 2001). Pioneer species employing a fast relative growth rate, low wood density ecological strategy would be benefitted by repeated setbacks to early successional conditions (Westoby 1998). Abundance weighted means may obscure the true pattern where differentiation in ecological strategy is strong, due to their inability to capture multimodality in trait distributions. In this case, greater abundance of these species would drive down mean wood density values through the upper ranges of disturbance intensity. This observation offers a potential explanation for the goodness of fit of quadratic models which begin to dip after reaching an apex at three quarters of their maximum value, rather than simply approaching an asymptote (e.g. Fig 3a,d, Fig 4a,b,d). It is difficult to substantiate this suggestion using our dataset, however, as only a few data points are present in the upper ranges of disturbance intensity.

There is another potential mechanism for coping with harsh environmental conditions, whereby some species have the ability to radically change their wood density throughout their life history. *Casuarina cunninghamiana*, for instance, is a rheophytic species (i.e. confined to frequently flooded substrates) whose entire life history revolves around response to flooding disturbance (van Steenis 1981). After seeding *en masse* onto fresh substrate within or close to the river channel (an ‘invader’ strategy), dense stands of flexible-stemmed saplings emerge, protecting each other from flood flows (Woolfrey & Ladd 2001). Self-thinning subsequently occurs, and our measurements from adult plants indicate that stem wood density increases during maturation; the plant has shifted towards the ‘resister’ phenotype. This change in strategy is likewise difficult to capture using abundance weighted means.

Under our argument, where hardy rheophytic species use high wood density ecological strategies to cope with powerful floods and unpredictable watering regimes, we would expect species such as *Casuarina cunninghamiana* and *Tristaniopsis laurina* to have the highest wood density in our dataset. However both species exhibit highly variable trait values, ranging approximately between the median value and the 75th percentile. As with *C. cunninghamiana,* *T. laurina* is a light-demanding coloniser of within and near-channel landforms (Webb, Erskine & Dragovich 2002). By establishing in close proximity to the channel, seedlings of these species must balance the risks of flooding with the advantages of growth unencumbered by competition for light or space. Maintaining a high relative growth rate, at least until the trees are physically large enough to endure flooding, allows these species to quickly fill space and build photosynthetic tissue (Melick 1990). If parallels with tropical rainforest species (King *et al.* 2006; Kraft, Valencia & Ackerly 2008; Poorter *et al.* 2008, 2010; Wright *et al.* 2010) hold, this strategy will not be conducive to setting down dense wood. In addition to morphological adaptations in *T. laurina* such as multi-stemmedness, narrow leaves and growth streamlined against the direction of flow (van Steenis 1981; Webb *et al.* 2002), the trade-off between flood resistance and rapid resource acquisition and growth during establishment serves to explain why the optimal wood density for rheophytic species might occupy a central position along the axis of wood density. Nonetheless, the wide plasticity in wood density shown by *C. cunninghamiana* and *T. laurina* suggests that intraspecific variation contributes to the species’ capacity to track hydrological gradients.

Notably, it is the tall, facultative riparian species from rainforest sites that had the highest wood density in our dataset.Lacking the morphological adaptations required to thrive directly along the channel edge, these species may rely solely on generating mechanically strong stems to withstand flooding. High wood density species tended to occur further up the bank, so would be subject to only the more intense flooding events. Since succession typically advances with elevation above the channel edge (Tabacchi, Correll & Hauer 1998), this observation agrees with previous studies showing increasing wood density along a successional gradient (Falster & Westoby 2005). Further parallels in the existing wood density literature are also evident here, where high wood density individuals were much less likely to experience major wind damage following a cyclone (Curran *et al.* 2008).

The gradient identified by principal components analysis integrates predictability of water availability, seasonality and flood intensity into a single axis of hydrological variation. It is not possible to tease out individual drivers of variation in wood density, as the conditions associated with both environmental unpredictability and mechanical disturbance act in unison to constrain community wood density to higher mean values. Based on our findings, hydrological regionalisation frameworks that distinguish between rivers according to predictability and perenniality of flow provide a basis for predicting wood density.

Hydrological classification therefore becomes useful in projecting changes to the functional attributes of riparian plant communities under altered flow conditions. In the south-eastern Australian context, changing flow conditions are caused by damming and water extraction, and the changing climate (Australian State of the Environment Committee 2011). Artificial flow modification by damming and water extraction reduces overall flow volume and the magnitude and frequency of high flow events, while increasing flow predictability, altering seasonality and limiting channel-floodplain connectivity (Maheshwari, Walker & McMahon 1995; Graf 2006). In these altered conditions, terrestrial species with softer wood and faster growth rates may encroach on what was once the province of rheophytic assemblages adapted to flooding and less predictable hydrological conditions. The converse of this situation is presented by predictions of future climatic conditions: in Australia, warming of 0.4 – 0.7oC has occurred since 1950 (Hennessy *et al.* 2007), associated with a reduction in rainfall across southern and eastern regions of the continent (Smith 2004), and an increase in intensity and frequency of droughts (Hennessy *et al.* 2008). Extreme rainfall events are predicted to become more prevalent, even in areas where the trend is towards mean reductions in annual or seasonal rainfall (Chiew *et al.* 2009).River discharge in Australia is known to be particularly sensitive to the El Nino-Southern Oscillation (ENSO) phenomenon that is an integral driver of the continent’s climate patterns (Nicholls 1989; Ward *et al.* 2010). Projected increases in climatic variability (Hennessy *et al.* 2008) may therefore overlay the already strong natural variability induced by ENSO to produce significant alterations to streamflow. Under such conditions, near-channel abundance of opportunistic, facultative riparian species may decline in favour of rheophytic species whose ecological strategies are optimized to harsh hydrological conditions. If changes in spatial extent of climate zones can be related to changes in runoff - a complicated, but progressing area of research in hydroclimatology (Peel & Bloschl 2011) – functional approaches to ecohydrology can give insight into likely changes in riparian plant assemblages and associated changes in ecosystem function.

Our study emphasises the importance of hydrological conditions, particularly disturbance and environmental unpredictability, as determinants of ecological strategy in riparian plant communities. This is likely to have significant ecological consequences for riparian plant communities in an Australian context, as well as in other regions where increasing climatic variability and frequency of extreme events are hallmarks of climate change predictions.

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**Data availability**

All data are available through DataDryad (datadryad.org). DOI to be provided upon publication.

**References**

Adler, P.B., Salguero-Gómez, R., Compagnoni, A., Hsu, J.S., Ray-Mukherjee, J., Mbeau-Ache, C. & Franco, M. (2014) Functional traits explain variation in plant life history strategies. *Proceedings of the National Academy of Sciences of the United States of America*, **111**, 740–5.

Arthington, A., Mackay, S., James, C., Rolls, R., Sternberg, D., A, B. & SJ, C. (2012) Ecological limits of hydrologic alteration: a test of the ELoHA framework in south-east Queensland. Waterlines Report Series No. 75. *National Water Commission, Canberra, Australia*.

Augspurger, C.K. & Kelly, C.K. (1984) Pathogen mortality of tropical tree seedlings: experimental studies of the effects of dispersal distance, seedling density, and light conditions. *Oecologia*, **61**, 211–217.

Australian State of the Environment Committee. (2011) Australia state of the environment 2011—in brief. Independent report to the Australian Government Minister for Sustainability, Environment, Water. *Population and Communities. Canberra: DSEWPaC*, 20–24.

Brierley, G.J. & Fryirs, K.A. (2005) *Geomorphology and River Management: Applications of the River Styles Framework*. John Wiley & Sons.

Catford, J.A., Naiman, R.J., Chambers, L.E., Roberts, J., Douglas, M. & Davies, P. (2012) Predicting Novel Riparian Ecosystems in a Changing Climate. *Ecosystems*, **16**, 382–400.

Chave, J., Coomes, D., Jansen, S., Lewis, S.L., Swenson, N.G. & Amy, E. (2009) Towards a worldwide wood economics spectrum. *Ecology Letters*, **12**, 351–366.

Chiew, F.H.S., Teng, J., Vaze, J., Post, D. a., Perraud, J.M., Kirono, D.G.C. & Viney, N.R. (2009) Estimating climate change impact on runoff across southeast Australia: Method, results, and implications of the modeling method. *Water Resources Research*, **45**, 1–17.

Coley, P. (1983) Herbivory and defensive characteristics of tree species in a lowland tropical forest. *Ecological monographs*, **53**, 209–234.

Corenblit, D., Steiger, J., Gurnell, A.M., Tabacchi, E. & Roques, L. (2009) Control of sediment dynamics by vegetation as a key function driving biogeomorphic succession within fluvial corridors. *Earth Surface Processes and Landforms*, **1810**, 1790–1810.

Cornelissen, J.H.C.A., Lavorel, S.B., Garnier, E.B., Díaz, S.C., Buchmann, N.D., Gurvich, D.E.C., Reich, P.B.E., Steege, H.F., Morgan, H.D.G., Van Der Heijden, M.G.A., Pausas, J.G.H. & Poorter, H.I. (2003) A handbook of protocols for standardised and easy measurement of plant functional traits worldwide. *Australian Journal of Botany*, **51**, 335–380.

Curran, T.J., Gersbach, L.N., Edwards, W. & Krockenberger, A.K. (2008) Wood density predicts plant damage and vegetative recovery rates caused by cyclone disturbance in tropical rainforest tree species of North Queensland, Australia. *Austral Ecology*, **33**, 442–450.

Falster, D.S. (2006) Sapling strength and safety: the importance of wood density in tropical forests. *The New phytologist*, **171**, 237–9.

Falster, D. & Westoby, M. (2005) Alternative height strategies among 45 dicot rain forest species from tropical Queensland, Australia. *Journal of Ecology*, **93**, 521–535.

Finlayson, B.L. & McMahon, T.A. (1988) Australia vs. the world: a comparative analysis of streamflow characteristics. *Fluvial Geomorphology of Australia*, 17–40.

Graf, W. (2006) Downstream hydrologic and geomorphic effects of large dams on American rivers. *Geomorphology*, **79**, 336–360.

Gutschick, V.P. & BassiriRad, H. (2003) Extreme events as shaping physiology, ecology, and evolution of plants: toward a unified definition and evaluation of their consequences. *New Phytologist*, **160**, 21–42.

Hacke, U.G., Sperry, J.S., Pockman, W.T., Davis, S.D. & McCulloh, K. a. (2001) Trends in wood density and structure are linked to prevention of xylem implosion by negative pressure. *Oecologia*, **126**, 457–461.

Healy, M., Raine, A., Parsons, L. & Cook, N. (2012) *River Condition Index in New South Wales: Method Development and Application. NSW Office of Water*. Sydney.

Hennessy, K., Fawcett, R., Kirono, D., Mpelasoka, M., Jones, D., Bathols, J., Whetton, P., Stafford Smith, M., Howden, M., Mitchell, C. & Plummer, N. (2008) *An Assessment of the Impact of Climate Change on the Nature and Frequency of Exceptional Climatic Events. Australian Government, Bureau of Meterology*. Department of Agriculture, Fisheries and Forestry, 2008., Canberra, Australia.

Hennessy, K., Fitzharris, B., Bates, B.C., Harvey, N., SM, H., L, H., J, S. & Warrick, R. (2007) *Climate Change 2007: Impacts, Adaptation and Vulnerability. Contribution of Working Group II to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change* (eds M. Parry, O. Canziani, J. Palutikof, P. van der Linden, and C. Hanson). Cambridge University Press, Cambridge.

Jacobsen, A.L., Agenbag, L., Esler, K.J., Pratt, R.B., Ewers, F.W. & Davis, S.D. (2007) Xylem density, biomechanics and anatomical traits correlate with water stress in 17 evergreen shrub species of the Mediterranean-type climate region of South Africa. *Journal of Ecology*, **95**, 171–183.

Jacobsen, A. & Ewers, F. (2005) Do xylem fibers affect vessel cavitation resistance? *Plant Physiology*, **139**, 546–556.

Kattge, J., Díaz, S., Lavorel, S., Prentice, I.C., Leadley, P., Bönisch, G., Garnier, E., Westoby, M., Reich, P.B., Wright, I.J., Cornelissen, J.H.C., Violle, C., Harrison, S.P., Van Bogedom, P.M., Reichstein, M., Enquist, B.J., Soudzilovskaia, N. a., Ackerly, D.D., Anand, M., Atkin, O., Bahn, M., Baker, T.R., Baldocchi, D., Bekker, R., Blanco, C.C., Blonder, B., Bond, W.J., Bradstock, R., Bunker, D.E., Casanoves, F., Cavender-Bares, J., Chambers, J.Q., Chapin Iii, F.S., Chave, J., Coomes, D., Cornwell, W.K., Craine, J.M., Dobrin, B.H., Duarte, L., Durka, W., Elser, J., Esser, G., Estiarte, M., Fagan, W.F., Fang, J., Fernández-Méndez, F., Fidelis, A., Finegan, B., Flores, O., Ford, H., Frank, D., Freschet, G.T., Fyllas, N.M., Gallagher, R. V., Green, W. a., Gutierrez, a. G., Hickler, T., Higgins, S.I., Hodgson, J.G., Jalili, A., Jansen, S., Joly, C. a., Kerkhoff, a. J., Kirkup, D., Kitajima, K., Kleyer, M., Klotz, S., Knops, J.M.H., Kramer, K., Kühn, I., Kurokawa, H., Laughlin, D., Lee, T.D., Leishman, M., Lens, F., Lenz, T., Lewis, S.L., Lloyd, J., Llusià, J., Louault, F., Ma, S., Mahecha, M.D., Manning, P., Massad, T., Medlyn, B.E., Messier, J., Moles, a. T., Müller, S.C., Nadrowski, K., Naeem, S., Niinemets, Ü., Nöllert, S., Nüske, A., Ogaya, R., Oleksyn, J., Onipchenko, V.G., Onoda, Y., Ordoñez, J., Overbeck, G., Ozinga, W. a., Patiño, S., Paula, S., Pausas, J.G., Peñuelas, J., Phillips, O.L., Pillar, V., Poorter, H., Poorter, L., Poschlod, P., Prinzing, A., Proulx, R., Rammig, A., Reinsch, S., Reu, B., Sack, L., Salgado-Negret, B., Sardans, J., Shiodera, S., Shipley, B., Siefert, A., Sosinski, E., Soussana, J.-F., Swaine, E., Swenson, N., Thompson, K., Thornton, P., Waldram, M., Weiher, E., White, M., White, S., Wright, S.J., Yguel, B., Zaehle, S., Zanne, a. E. & Wirth, C. (2011) TRY - a global database of plant traits. *Global Change Biology*, **17**, 2905–2935.

Kennard, M.J., Pusey, B.J., Olden, J.D., Mackay, S.J., Stein, J.L. & Marsh, N. (2010) Classification of natural flow regimes in Australia to support environmental flow management. *Freshwater Biology*, **55**, 171–193.

King, D.A., Davies, S.J., Tan, S. & Noor, N.S.M. (2006) The role of wood density and stem support costs in the growth and mortality of tropical trees. *Journal of Ecology*, **94**, 670–680.

Kraft, N.J.B., Metz, M.R., Condit, R.S. & Chave, J. (2010) The relationship between wood density and mortality in a global tropical forest data set. *The New Phytologist*, **188**, 1124–36.

Kraft, N., Valencia, R. & Ackerly, D. (2008) Functional traits and niche-based tree community assembly in an Amazonian forest. *Science*, 580–582.

Lens, F., Sperry, J.S., Christman, M.A., Choat, B., Rabaey, D. & Jansen, S. (2011) Testing hypotheses that link wood anatomy to cavitation resistance and hydraulic conductivity in the genus Acer. *The New Phytologist*, **190**, 709–23.

Maheshwari, B., Walker, K. & McMahon, T. (1995) Effects of regulation on the flow regime of the River Murray, Australia. *Regulated Rivers: Research and Management*, **10**, 15–38.

Marsh, N.A., Stewardson, M.J. & Kennard, M.J. (2003) River analysis package. Cooperative Research Centre for Catchment Hydrology, Monash University Melbourne. Version 1. *Software Version*, **1**.

Martínez-Cabrera, H.I., Jones, C.S., Espino, S. & Schenk, H.J. (2009) Wood anatomy and wood density in shrubs: Responses to varying aridity along transcontinental transects. *American Journal of Botany*, **96**, 1388–98.

McGill, B.J., Enquist, B.J., Weiher, E. & Westoby, M. (2006) Rebuilding community ecology from functional traits. *Trends in Ecology & Evolution*, **21**, 178–85.

Melick, D. (1990) Regenerative succession of Tristaniopsis laurina and Acmena smithii in riparian warm temperate rain-forest in Victoria, in relation to light and nutrient regimes. *Australian Journal of Botany*, **38**, 111–120.

Merritt, D.M., Scott, M.L., Poff, N.L., Auble, G.T. & Lytle, D. a. (2010) Theory, methods and tools for determining environmental flows for riparian vegetation: riparian vegetation-flow response guilds. *Freshwater Biology*, **55**, 206–225.

Naiman, R. & Decamps, H. (1997) The ecology of interfaces: riparian zones. *Annual Review of Ecology and Systematics*, **28**, 621–658.

Nicholls, N. (1989) Sea surface temperatures and Australian winter rainfall. *Journal of Climate*, **2**, 965–973.

Niklas, K.J. & Spatz, H.-C. (2010) Worldwide correlations of mechanical properties and green wood density. *American Journal of Botany*, **97**, 1587–94.

Oksanen, J., Blanchet, F.G., Kindt, R., Legendre, P., Minchin, P.R., O’Hara, R.B., Simpson, G.L., Solymos, P., Stevens, M.H.H. & Wagner, H. (2013) vegan: Community Ecology Package.

Olden, J.D. & Poff, N.L. (2003) Redundancy and the choice of hydrologic indices for characterizing streamflow regimes. *River Research and Applications*, **19**, 101–121.

Peel, M.C. & Bloschl, G. (2011) Hydrological modelling in a changing world. *Progress in Physical Geography*, **35**, 249–261.

Peel, M., McMahon, T. & Finlayson, B. (2004) Continental differences in the variability of annual runoff-update and reassessment. *Journal of Hydrology*, **295**, 185–197.

Poff, N.L., Richter, B.D., Arthington, A.H., Bunn, S.E., Naiman, R.J., Kendy, E., Acreman, M., Apse, C., Bledsoe, B.P., Freeman, M.C., Henriksen, J., Jacobson, R.B., Kennen, J.G., Merritt, D.M., O’™Keeffe, J.H., Olden, J.D., Rogers, K., Tharme, R.E. & Warner, A. (2010) The ecological limits of hydrologic alteration (ELOHA): a new framework for developing regional environmental flow standards. *Freshwater Biology*, **55**, 147–170.

Poorter, L., McDonald, I., Alarcón, A., Fichtler, E., Licona, J.-C., Peña-Claros, M., Sterck, F., Villegas, Z. & Sass-Klaassen, U. (2010) The importance of wood traits and hydraulic conductance for the performance and life history strategies of 42 rainforest tree species. *The New Phytologist*, **185**, 481–92.

Poorter, L., Wright, S.J., Paz, H., Ackerly, D.D., Condit, R., Ibarra-Manríquez, G., Harms, K.E., Licona, J.C., Martínez-Ramos, M., Mazer, S.J., Muller-Landau, H.C., Peña-Claros, M., Webb, C.O. & Wright, I.J. (2008) Are functional traits good predictors of demographic rates? Evidence from five neotropical forests. *Ecology*, **89**, 1908–20.

Preston, K.A., Cornwell, W.K. & Denoyer, J.L. (2006) Wood density and vessel traits as distinct correlates of ecological strategy in 51 California coast range angiosperms. *The New Phytologist*, **170**, 807–18.

R Core Team. (2013) R: A Language and Environment for Statistical Computing.

Reich, P. & Wright, I. (2003) The evolution of plant functional variation: traits, spectra, and strategies. *International Journal of Plant Sciences*, **164**, 146–164.

Russo, S.E., Jenkins, K.L., Wiser, S.K., Uriarte, M., Duncan, R.P. & Coomes, D.A. (2010) Interspecific relationships among growth, mortality and xylem traits of woody species from New Zealand. *Functional Ecology*, **24**, 253–262.

Smith, I. (2004) An assessment of recent trends in Australian rainfall. *Australian Meteorological Magazine*, **53**, 163–173.

Van Steenis, C.G.G.J. (1981) *Rheophytes of the World: An Account of the Flood-Resistant Flowering Plants and Ferns and the Theory of Autonomous Evolution*. Sijthoff & Noordhoff Alphen aan den Rijn, Netherlands.

Swenson, N. & Enquist, B. (2007) Ecological and evolutionary determinants of a key plant functional trait: wood density and its community-wide variation across latitude and elevation. *American Journal of Botany*, **94**, 451–459.

Tabacchi, E., Correll, D. & Hauer, R. (1998) Development, maintenance and role of riparian vegetation in the river landscape. *Freshwater …*, **40**, 497–516.

Valladares, F., Balaguer, L., Martinez-Ferri, E., Perez-Corona, E. & Manrique, E. (2002a) Plasticity, instability and canalization: is the phenotypic variation in seedlings of sclerophyll oaks consistent with the environmental unpredictability of Mediterranean ecosystems? *New Phytologist*, **156**, 457–467.

Valladares, F., Chico, J. & Aranda, I. (2002b) The greater seedling high-light tolerance of Quercus robur over Fagus sylvatica is linked to a greater physiological plasticity. *Trees*, **16**, 395–403.

Valladares, F., Gianoli, E. & Gómez, J. (2007) Ecological limits to plant phenotypic plasticity. *New Phytologist*, **176**, 749–63.

Ward, P.J., Beets, W., Bouwer, L.M., Aerts, J.C.J.H. & Renssen, H. (2010) Sensitivity of river discharge to ENSO. *Geophysical Research Letters*, **37**.

Webb, A., Erskine, W. & Dragovich, D. (2002) Flood-driven formation and destruction of a forested flood plain and in-channel benches on a bedrock-confined stream: Wheeny Creek, southeast Australia. *nternational Association of Hydrological Sciences*, **276**, 203–210.

Weimann, M. & Williamson, G. (2002) Geographic variation in wood specific gravity: effects of latitude, temperature and precipitation. *Wood and Fiber Science*, **34**, 96–107.

Westoby, M. (1998) A leaf-height-seed (LHS) plant ecology strategy scheme. *Plant and Soil*, **199**, 213–227.

Westoby, M., Falster, D.S., Moles, A.T., Vesk, P. a. & Wright, I.J. (2002) PLANT ECOLOGICAL STRATEGIES: Some Leading Dimensions of Variation Between Species. *Annual Review of Ecology and Systematics*, **33**, 125–159.

Westoby, M. & Wright, I.J. (2006) Land-plant ecology on the basis of functional traits. *Trends in Ecology & Evolution*, **21**, 261–8.

Woolfrey, A.R. & Ladd, P.. (2001) Habitat preference and reproductive traits of a major Australian riparian tree species (Casuarina cunninghamiana). *Australian Journal of Botany*, **49**, 705–715.

Wright, S.J., Kitajima, K., Kraft, N.J.B., Reich, P.B., Wright, I.J., Bunker, D.E., Condit, R., Dalling, J.W., Davies, S.J., Díaz, S., Engelbrecht, B.M.J., Harms, K.E., Hubbell, S.P., Marks, C.O., Ruiz-Jaen, M.C., Salvador, C.M. & Zanne, A.E. (2010) Functional traits and the growth-mortality trade-off in tropical trees. *Ecology*, **91**, 3664–74.

Zieminska, K., Butler, D.W., Gleason, S.M., Wright, I.J. & Westoby, M. (2013) Fibre wall and lumen fractions drive wood density variation across 24 Australian angiosperms. *AoB PLANTS*, **5**, plt046.

**Table 1.** Locations, characteristics and sampling of field sites.

|  |  |  |  |  |
| --- | --- | --- | --- | --- |
| **Site** | **Longitude** | **Latitude** | **Hydrological class** | **Number of species sampled** |
| Snowy Creek | 147.413 | -36.569 | stable winter baseflow | 3 |
| Gibbo River | 147.709 | -36.756 | stable winter baseflow | 4 |
| Goodradigbee River | 147.826 | -36.444 | stable winter baseflow | 7 |
| Nariel Creek | 148.731 | -36.421 | stable winter baseflow | 4 |
| Jacob’s River | 148.427 | -36.727 | stable winter baseflow | 4 |
| Tuross River at Belowra | 149.709 | -36.201 | unpredictable baseflow | 4 |
| Genoa River | 149.321 | -37.174 | unpredictable baseflow | 5 |
| Wallagaraugh River | 149.714 | -37.371 | unpredictable baseflow | 1 |
| Mann River | 152.105 | -29.695 | unpredictable baseflow | 4 |
| Cataract Creek | 152.217 | -28.934 | unpredictable baseflow | 3 |
| Jilliby Creek | 151.389 | -33.246 | unpredictable intermittent | 5 |
| Sportsmans Creek | 142.981 | -29.467 | unpredictable intermittent | 8 |
| Mammy Johnsons River | 151.979 | -32.244 | unpredictable intermittent | 7 |
| Wadbilliga River | 149.694 | -36.259 | unpredictable intermittent | 4 |
| Tuross River downstream of Wadbilliga junction | 149.761 | -36.197 | unpredictable intermittent | 6 |

**Table 2.** Flow characteristics of the three river hydrological classes used in this study. The reader is directed to Kennard et al. (2010) for the complete characterisation and derivation of these classes. It is worthwhile to note here that these three classes span roughly half of the range of hydrological variability within the Australian continent. The extreme hydrological variability within the arid and semi-arid regions that dominate the centre of the continent (Finlayson & McMahon 1988) is not represented here.

|  |  |  |  |
| --- | --- | --- | --- |
|  | ***Stable winter baseflow*** | ***Unpredictable baseflow*** | ***Unpredictable intermittent*** |
| **Geographic distribution** | Restricted to southern half of Australia, primarily South East Coast and South West Coast drainage divisions and Tasmania. | Distributed widely across southern and eastern Australia. | Distributed widely along east coast of Australia. |
| **Perenniality** | Perennial | Perennial | Intermittent |
| **Seasonality** | Strongly winter dominated | Weak seasonal signal | Highly unpredictable |
| **Flow variability** | Low | Moderate | High |
| **Flood characteristics** | Small, frequent, short duration. Slow rise and fall rates. | Roughly 50% higher magnitude than stable winter baseflow streams, with similar duration and faster rise and fall rates. | Higher magnitude than unpredictable baseflow streams, with similar duration and faster rise and fall rates. |
| **Baseflow contribution** | High | High, but lower than stable winter baseflow | Intermediate |

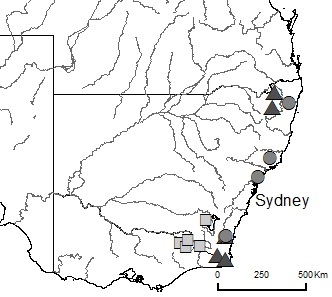
**Table 3.** Hydrological parameters used as metrics of variability in high flow magnitude and frequency and predictability and consistency of water availability in the riparian environment. \* - normalised by mean daily flow (ML/day)

|  |  |  |  |  |
| --- | --- | --- | --- | --- |
| **Parameter** | **Abbreviation** | | **Units** | **Description** |
| *Flood frequency and magnitude* | | | | |
| Mean magnitude of high spells\* | HSPeaknorm | dimensionless | | High spells are periods of flow above the 95th percentile on the flow duration curve. We were interested in how frequently these conditions occurred over the time series as well as the mean magnitude of peak flows during these periods. 20 year average return interval (ARI) floods are extreme flow events that have the potential to re-work the fluvial landscape. Together, these metrics indicate the intensity and frequency of mechanical stress experienced by plants in the riparian zone. |
| CV of all years’ mean high spell magnitude | CVAnnHSPeak | dimensionless | |
| 20 year ARI flood magnitude\* | AS20YrARInorm | dimensionless | |
| Mean of all years’ number of high spells | MDFAnnHSNum | year-1 | |
| CV of all years’ number of high spells | CVAnnHSNum | dimensionless | |
| *Rise and fall rates* | | | | |
| Mean rate of rise \* | MRateRisenorm | day-1 | | Rise and fall rates represent flow ‘flashiness’. Fast rise rates are associated with flood waves and intense mechanical stress to plant stems. Slow fall rates keep exposed substrate moist for longer periods, which may produce favourable conditions for germination. Historical discharge records are unfortunately limited to daily resolution, so are unable to fully capture flood discharge shapes. High variability between years indicates the occurrence of extreme events which may not have been captured by the mean value. |
| Mean rate of fall \* | MRateFallnorm | day-1 | |
| CV of all years’ mean rate of rise | CVAnnMRateRise | dimensionless | |
| CV of all years’ mean rate of fall | CVAnnMRateFall | dimensionless | |
| *Baseflow index* | | | | |
| Baseflow index | BFI | dimensionless | | Baseflow index is calculated using the ratio of flow during average conditions to total flow. It is a useful metric of consistency of water availability, in that it is maximised when average flow conditions dominate, and minimised when total flow is dominated by above average flow events. Intra-annual variability in baseflow index measures how predictable baseflow index is between years. |
| CV of all years’ Baseflow Index | CVAnnBFI | dimensionless | |
| *Low flow magnitude, frequency and duration* | | | | |
| CV of all years’ mean low spell magnitude | LSPeaknorm | dimensionless | | Low spells are periods of flow below the 5th percentile on the flow duration curve. We were interested in how frequently these conditions occurred over the time series as well as the mean and interannual variability in magnitude and duration of low flows. |
| Mean magnitude of low spells | CVAnnLSPeak | dimensionless | |
| Mean of all years’ number of low spells | MDFAnnLSNum | year-1 | |
| CV of all years’ number of low spells | CVAnnLSNum | dimensionless | |
| Mean duration of low spells | LSMeanDur | days | |
| CV of all years’ low spell mean duration | CVAnnLSMeanDur | dimensionless | |
| Mean flow during driest week of the year\* | MA.7daysMinMeannorm | dimensionless | |
| Mean days per year under 0.1ML/day flow | MDFAnnUnder0.1 | days/year | |
| CV of all years’ days per year under 0.1ML/day flow | CVAnnMDFAnnUnder0.1 | dimensionless | |
| *Colwell’s indices* | | | | |
| Constancy of monthly mean daily flow | C\_MDFM | dimensionless | | Colwell’s indices provide a measure of the seasonal predictability of flow events and therefore water availability within the riparian zone. Constancy (C) measures uniformity of flow across seasons, and is maximised when flow conditions do not differ between seasons. Contingency (M) is a measure of interannual uniformity in seasonal flow patterns, and is maximized when seasonal patterns of flow are consistent between years.  We generated Colwell’s indices for both average flow conditions and minimum flows conditions. |
| Contingency of monthly mean daily flow | M\_MDFM | dimensionless | |
| Constancy based on monthly minimum daily flow | C\_MinM | dimensionless | |
| Contingency based on monthly minimum daily flow | M\_MinM | dimensionless | |

**Table 4.** Statistics for regression models comparing hydrological metrics with site mean wood density. p.adj denotes p values adjusted by the Benjamini-Hochberg method. Statistics for models where Snowy Creek was removed as an outlier are also given. Significant results are shown in italics. The initial best fit for AS20YrARInorm was an exponential model, but after removal of Snowy Creek, values are given for a quadratic model which provided a better fit. The model for MA.7daysMinMeannorm was made non-significant after p-value adjustment, but returned to significance following outlier removal. CVAnnHSPeak was non-significant initially but a significant relationship became apparent following outlier removal.

|  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- |
|  | All values included | |  | Snowy Creek value removed | | |
| *Metric* | *p* | *p.adj* | *R2* | *p* | *p.adj* | *R2* |
| M\_MinM | *0.0038* | *0.0432* | *0.552* | *0.0002* | *0.0008* | *0.740* |
| CVAnnBFI | *0.0064* | *0.0432* | *0.557* | *0.0001* | *0.0007* | *0.838* |
| CVAnnMRateRise | *0.0068* | *0.0432* | *0.515* | *0.0001* | *0.0007* | *0.787* |
| M\_MDFM | *0.0094* | *0.0432* | *0.443* | *0.0018* | *0.0039* | *0.606* |
| C\_MDFM | *0.0098* | *0.0432* | *0.450* | *0.0001* | *0.0007* | *0.758* |
| CVAnnMRateFall | *0.0117* | *0.0432* | *0.434* | *0.0001* | *0.0007* | *0.783* |
| AS20YrARInorm | *0.0124* | *0.0279* | *0.393* | *0.0057* | *0.0114* | *0.668* |
| LSPeaknorm | *0.0128* | *0.0432* | *0.427* | *0.0002* | *0.0008* | *0.724* |
| HSPeaknorm | *0.0144* | *0.0432* | *0.415* | *0.0005* | *0.0013* | *0.708* |
| BFI | *0.0180* | *0.0480* | *0.434* | *0.0001* | *0.0007* | *0.816* |
| MA.7daysMinMeannorm | *0.0355* | 0.0852 | 0.328 | *0.0003* | *0.0009* | *0.755* |
| CVAnnHSPeak | 0.0751 | 0.1502 | 0.293 | *0.0017* | *0.0039* | *0.754* |
| MRateRisenorm | 0.1631 | 0.2899 | 0.348 | 0.0730 | 0.1348 | 0.528 |
| CVAnnHSNum | 0.1691 | 0.2899 | 0.164 | 0.2300 | 0.3450 | 0.135 |
| MDFAnnLSNum | 0.1908 | 0.3053 | 0.208 | 0.2600 | 0.3671 | 0.183 |
| MRateFallnorm | 0.2098 | 0.3061 | 0.283 | 0.1100 | 0.1886 | 0.443 |
| CVAnnLSPeak | 0.2168 | 0.3061 | 0.245 | 0.1500 | 0.2400 | 0.339 |
| LSMeanDur | 0.4115 | 0.5487 | 0.180 | 0.4300 | 0.5280 | 0.159 |
| CVAnnLSNum | 0.4417 | 0.5579 | 0.052 | 0.4400 | 0.5280 | 0.054 |
| C\_MinM | 0.4919 | 0.5903 | 0.259 | 0.3700 | 0.4933 | 0.427 |
| MDFAnnUnder0.1 | 0.5904 | 0.6747 | 0.071 | 0.5100 | 0.5829 | 0.105 |
| MDFAnnZer | 0.6360 | 0.6938 | 0.092 | 0.5600 | 0.6109 | 0.130 |
| MDFAnnHSNum | 0.6885 | 0.7184 | 0.262 | 0.8100 | 0.8100 | 0.254 |
| CVAnnLSMeanDur | 0.8483 | 0.8483 | 0.029 | 0.6700 | 0.6991 | 0.028 |

**Figures**

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***Figure 1.*** *Location of fifteen field study sites across south-eastern Australia chosen to represent the three major hydrological classes of south-east Australia. Hydrological class membership is denoted by: • stable winter baseflow, ▲ unpredictable baseflow, ■ unpredictable intermittent. Note that the points representing the two southern-most unpredictable baseflow sites are overlapping.*



**Figure 2.** Comparison of mean wood density between hydrological classes using a) abundance weighted means, b) means of raw wood density values. Error bars represent standard error of the mean. Hydrological classes are: (1) stable winter baseflow, (2) unpredictable intermittent.

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**Figure 3.** Relationships between abundance weighted mean wood density and hydrological metrics describing a) interannual variability in flood rise rates (CVAnnMRateRise), b) interannual variability in flood fall rates (CVAnnMRateFall), c) mean high flow magnitude (HSPeaknorm), d) interannual variability in high flow magnitude (CVAnnHSPeak), e) magnitude of the 20 year average return interval flood (AS20YrARInorm). Fitted lines depict ordinary least squares regression models. a. – d. are quadratic fits, e. is an exponential fit. Shaded areas depict the smoothed 95% confidence interval around the regression model. Hydrological class membership is denoted by: • stable winter baseflow, ▲ unpredictable baseflow, ■ unpredictable intermittent.

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**Figure 4.** Relationships between abundance weighted mean wood density and hydrological metrics describing a.) baseflow index (BFI), b.) interannual variability in baseflow index (CVAnnBFI), c.) contingency of monthly mean daily flow (M\_MDFM), d.) contingency of monthly mean daily flow (C\_MDFM), e.) constancy of monthly minimum daily flow (M\_MinM), f.) mean low flow magnitude (LSPeaknorm), g.) Mean flow during driest week of the year (MA.7days.MinMeannorm). Shaded areas depict the 95% confidence interval around the regression model. Hydrological class membership is denoted by: • stable winter baseflow, ▲ unpredictable baseflow, ■ unpredictable intermittent.

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**Figure 5.** Biplot of sites ordinated across the first two principal components (PC) of the PCA. Points represent positions of individual sites. Hydrological class membership is denoted by: • stable winter baseflow, ▲ unpredictable baseflow, ■ unpredictable intermittent. Arrows represent loadings of hydrological metrics across each PC.

**Table 1.** Locations, characteristics and sampling of field sites.

**Table 2.** Flow characteristics of the three river hydrological classes used in this study. The reader is directed to Kennard et al. (2010) for the complete characterisation and derivation of these classes. It is worthwhile to note here that these three classes span roughly half of the range of hydrological variability within the Australian continent. The extreme hydrological variability within the arid and semi-arid regions that dominate the centre of the continent (Finlayson & McMahon 1988) is not represented here.

**Table 3.** Hydrological parameters used as metrics of variability in high flow magnitude and frequency and predictability and consistency of water availability in the riparian environment. \* - normalised by mean daily flow (ML/day)

**Table 4.** Statistics for regression models comparing hydrological metrics with site mean wood density. p.adj denotes p values adjusted by the Benjamini-Hochberg method. Statistics for models where Snowy Creek was removed as an outlier are also given. Significant results are shown in italics. The initial best fit for AS20YrARInorm was an exponential model, but after removal of Snowy Creek, values are given for a quadratic model which provided a better fit. The model for MA.7daysMinMeannorm was made non-significant after p-value adjustment, but returned to significance following outlier removal. CVAnnHSPeak was non-significant initially but a significant relationship became apparent following outlier removal.

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